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Electrochemical Studies of Reactive Orange 12 on Glassy Carbon Electrode

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ABSTRACT

The cyclic voltammetric studies of reactive orange 12 (RO 12) was carried out in aqueous solution on glassy carbon electrode (GCE). The effect of pH on the electrochemical behaviour of reactive orange 12 was performed from pH 1.0 to 13.0 at scan rate 50 mV/s. The maximum peak current response was found in pH 4.0 others little bit less. At all pHs three anodic peaks in forward scan and two cathodic peak in the reverse scan were observed. The dye solution of RO 12 was exhibits oxidation and reduction peaks potentials at around -190, 030, 650 mV and -220, -6750 mV respectively vs Ag/AgCl The voltammogram of. A systematic study of the experimental parameters that affect the anodic differential pulse stripping voltammetric response was carried out. Maximum peak current conditions were arrived. Calibration curve was made under the maximum peak current conditions at different concentration RO 12. The concentration range studied for the determination of 300 to 700 ppb. The lower limit of detection is 200 ppb for GCE and the RSD 3.2%. The suitability of this method for the quantisation of dye in textile industries effluents was also ascertained.

KEY WORDS: Cyclic voltammetry, Reactive Orange 12, Stripping voltammetry, Glassy Carbon Electrode

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INTRODUCTION

The textile industry is one of the most polluting industries in terms of discharge volume and effluent composition¹. Reactive dyes are extensively used in the textile industry due to the presence of reactive groups in their molecules, which bind to fibers through covalent bonds. The main problem associated with the use of reactive dyes is their low affinity for fabrics resulting in relatively high dye concentrations in textile wastewater². Reactive dyes exhibit a wide range of different chemical structures, primarily based on substituted aromatic and heterocyclic groups. Since reactive dyes are highly soluble in water, their removal from wastewater is difficult by conventional treatment processes^{3, 4, 5}. A moderate difference was discussed in our preliminary study of the morphology of graphite anode electrode before and after the electrochemical process⁶. In the electrochemical oxidation process anode reaction, the substrate oxidation involves the transfer of one or more electrons from the substrate to the anode—direct electron transfer (DET). However, substrate oxidation rarely occurs without simultaneous oxidation of water, especially under amperostatic conditions. This opens up the possibility that attack on the substrate is initiated by reactive intermediates of water oxidation, such as hydroxyl radicals, hydrogen peroxide, hydroperoxy radicals and their basic conjugates, either free in solution or bound to the anode surface⁷.

Ceria-based composites have been previously developed as functional electrolytes for high performance of solid oxide fuel cells that require high functional electrolyte materials that can provide high ion conductivity for sufficient current output. These composites display hybrid proton and oxygen ion conduction. We developed further composite electrolyte materials containing a catalyst such as rare earth elements; gadolinium (Gd^{3+}), neodymium (Nd^{3+}) and samarium (Sm^{3+}) doped ceria ($Ce_{0.8}Gd_{0.2}O_2$, $Ce_{0.8}Nd_{0.2}O_2$ and $Ce_{0.8}Sm_{0.2}O_2$) to investigate the electrochemical treatment of C.I. Reactive Orange 107⁸. The discharge of dyes is a worldwide environmental problem. These dyes are widely applied in industries namely dyestuff, rubber, paper, leather, textiles, plastics and cosmetics^{9, 10, 11}. Reactive azo dyes are highly soluble in water and differ from all other categories of dyes in which they bind to the textile fibers namely cotton, wool and silk to form covalent bonds¹².

The effect of pH on the adsorption of the reactive orange 12 dye by rice husk was investigated at ambient temperature (25 °C). pH is one of the most significant parameters controlling the adsorption of dye by adsorbent particles, which changes both the surface charge of the adsorbents and the degree of ionization of different adsorbates¹³. Huge amounts of synthetic dyes are used in many industries, such as dyestuffs, leather, tanning, and textiles. In the textile industry, it was

estimated that approximately 10%–15% of total dyes utilized were lost during the dyeing processes and released into wastewater¹⁴. More than 60% of dyestuffs manufactured around the world are azo dyes¹⁵, which are recalcitrant, non biodegradable, carcinogenic, mutagenic, and toxic¹⁶. Wastewaters that include azo dyes must be effectively treated to prevent possible harm to aquatic life, since they may contain different pollutants such as dyestuff with various structures and textile auxiliary materials such as surface active materials and polyvinyl alcohol utilized in sizing processes¹⁷. The color removal efficiencies were 85% by ferrate (VI) method and 91% by sono-ferrate (VI) method. Kinetic studies were also performed for the decolonization of RO 16 under the optimized conditions at room temperature¹⁸. pH has the significant impact on absorption process and affected absorption capacity, dye solubility¹⁹, solution chemistry, and a surface of absorbent pore²⁰. The effect of pH on the adsorption capability was investigated between pH 2.0 and 12 at 30°C and the shaking rate was set at 150 rpm. The percentage removal of RO16 increased from 95% to 98.12% with an increase of the pH from 2 to 8²¹. The study was to develop electro-analytical strategies based on certain voltammetric techniques for the sensitive determination of RO12 in aqueous solution on glassy carbon electrode. Direct redox behavior of reactive orange 12 on GCE and as consequence, their determination was investigated by cyclic voltammetry (CV) and differential-pulsed voltammetry (DPV) techniques-based procedures were tested, developed and optimized for selective and detection of reactive orange 12.

EXPERIMENTAL METHODS

All reagents were of AR grade purchased commercially. The stock solution of reactive orange 12 was prepared by dissolving the substances in double distilled water purified from SG purification system. The supporting electrolyte (pHs) solutions were also prepared by same water. The pH values of the buffer solutions were measured by Hanna HI 2211 pH/ORP meter.

The electrochemical studies were carried out in exploratory and determination mode on a software connected CH Instruments Electrochemical Workstation (model CH 650C). The cell was made of glass, having a capacity of 15 mL and the Teflon made cell top was comprised of three separate holes for the insertion of electrodes viz. working electrode (Glassy Carbon), counter electrode (platinum foil) and reference electrode (Saturated calomel). The cell setup is kept in a CH Instrument Picoamp Booster and Faraday cage. The cell top also has the purging and blanketing facilities of nitrogen gas with separate tubes to remove oxygen gas. This setup enables to maintain an inert atmosphere above the sample solution throughout the experiment.

The surface of the glassy carbon electrode is apt to be contaminated by surface-active substances and other compounds. Thus the effective surface area and other properties may be

changed. Hence, it is necessary to treat the electrode before use, if a good reproducibility is required. Well-known and accepted procedures for the treatment of glassy carbon electrode were adopted. First the electrode was washed with water-ethanol-ammonia mixture followed by ethanol-ammonia, ethanol-acetic acid and ethyl acetate-ethanol mixture. It was then washed with distilled water and trichloro ethylene. With this method, oxides and surface-active substances and organic compounds were removed from the electrode surface. When electrode surface seriously contaminated, the most effective and, at the same time the simplest way employed for the renewal of the surface was to remove a layer of it by rubbing with 0.05 μM of fine powder of $\gamma\text{-Al}_2\text{O}_3$. After rubbing, the electrode was wiped with a filter paper and then rinsed with water. The fine particles of alumina adsorbed on glassy carbon electrode were removed by ultrasonication in the presence of water. Then, the electrode was thoroughly degreased with trichloroethylene and washed with deionized water and acetone. Glassy carbon electrode (GCE) surface was well developed for determination of trace amount of dye. The dye adsorbed on GCE is now a day's very useful for determination of ultra trace materials.

Table 1: Studied range and optimum experimental conditions in DPSV

Parameters	Range studied	Optimum value
Ph	1-13	4.0
Accumulation potential (mV)	-800 to 800	400
Accumulation time (Sec)	10-90	60
Initial scan potential (mV)	-1300 to -600	-1000
Pulse Height (PH) (mV)	25 to 200	150
Pulse width (PW) mSec	25 to 200	75
Scan Increment (SI) mV	2 to 20	16
Scan rate (SR) mV/sec	10 to 100	50
Stirring rate (rpm)	50 to 250	250
Rest period (Sec)	2 to 10	5

RESULTS AND DISCUSSION

Cyclic Voltammetric Behaviour

Cyclic voltammograms of RO 12 on GCE in acid, neutral and alkaline media, at different scan rates from 25 to 500 mV/s were recorded. There was significant response when the studies were carried out between pH 1.0 to 13.0. The background current was recorded for all scan rates and

subtracted in the appropriate potential range. The peak potential E_p (in milli volts) and peak current i_p (in microamperes) for respective concentrations and scan rates were measured. The data are obtained and discussed.

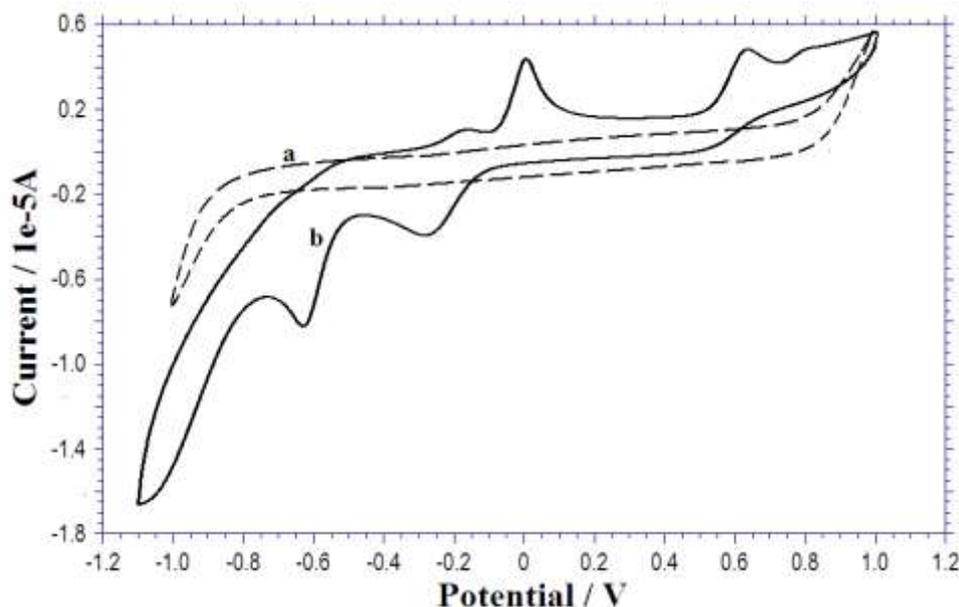


Figure 1, Cyclic voltammetric behavior of 300 ppm concentration of RO 12 on GCE in pH 4.0 and at 50 mV/s

Effect of pH

Since there was characteristic effect from pH 1.0, the cyclic voltammetric studies of RO 12 were carried out in the pH range 1.0 to 13.0. At all pHs three anodic peaks and two cathodic peak (Fig.1) in the reverse scan were observed. Of the three peaks, anodic peak III showed prominence with higher current in the pH range 1.0 to 13.0. The anodic peak III and cathodic peak lost their sharpness when the pH was increased. Hence the anodic peak III was considered for further studies to probe the oxidation mechanism. For the study of influence of pH on the cyclic voltammetric response, at pH 4 were chosen and the cyclic voltammograms were recorded for 300 ppm RO 12 at a scan rate of 100 mV/s. The peak currents and peak potentials were plotted against pH and the plots are given in figures 2. The peak current decreased with increase in pHs. The maximum peak current response was found in pH 4.0. This is due to faster electron transfer at acid pH 4.0 and indicates that the electrochemical oxidation of RO 12 is facilitated in acidic media. Hence it can be considered as an optimum pH for the study of RO 12. The peak potential also decreased with increase in pHs. The peak potential of anodic peak III shifted anodically with pH whereas that of cathodic peak shifted cathodically with pHs.

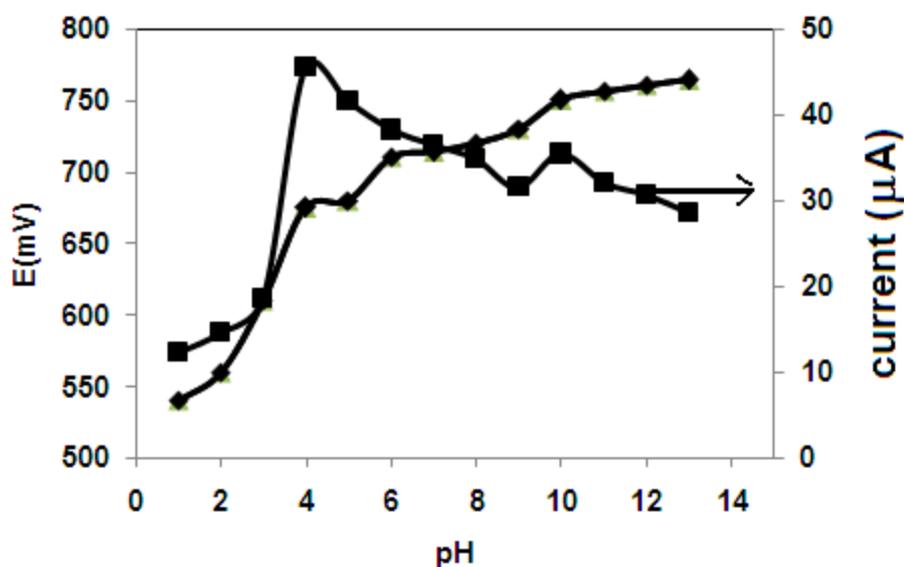


Figure 2, Effect of pH vs. peak current, potential curve on GCE

Effect of Scan Rate and Concentration

The scan rate was varied from 25 to 500 mV/s at the selected pH medium and the cyclic voltammetric behaviour of 300 ppm RO 12 was understood. Peak current was correlated with scan rate and square root of scan rate. The plot of log current versus log scan rate was also made. At constant scan rate 100 mV/s, the concentration of the substrate was varied and the effect on peak current was observed. The plot of current versus concentration of RO 12 was also made.

An aqueous solution of pH 4.0 was chosen as the medium for the study. Fig.1 represents the cyclic voltammogram of RO 12 recorded at pH 4.0. Three anodic peaks were observed at the potentials around -190, 030, 650 mV. In the reverse scan two cathodic peak was observed around at -220 and -6750 mV. The effect of scan rate was studied for the anodic peak I at a scan rate from 25 to 500 mV/s. The peak current was plotted against scan rate and resulted in a slightly curved line whereas the peak current increased linearly with square root of scan rate and resulted in a straight line with good correlation ($R^2 = 0.9902$). The log i_p vs. log v plot yielded a straight line with slope 0.1002. These facts revealed the nature of electrode reaction as diffusion controlled. The variation of peak potential with log scan rate resulted in a straight line ($\alpha n = 0.8855$). The potential difference between the anodic peak I and the cathodic peak was found to be around 100 mV. Hence, it can be considered that the oxidation of RO 12 at pH 4.0 may be considered as quasi reversible. The effect of concentration on peak current at a constant scan rate 100 mV/s was carried out by changing the concentration from 300 to 700 ppm. Increase in the concentration of RO 12 showed increase in the peak current as well as increase in peak potential. The plot of i_p vs. concentration yielded a straight

line. Hence, it may be concluded that at pH 4.0, irreversible diffusion controlled oxidation and reduction reactions are taking place apart from another oxidation at higher anodic potential.

Chronocoulometry

Diffusion coefficient is required for most electrochemical studies since the current response at an electrode depends on the rate of mass transport of the electroactive species to the electrode surface. The diffusion coefficient value of the compound was calculated from the slope of Anson's plot, if other parameters are known. In this case, the geometric area of the glassy carbon electrode used is 0.0314cm^2 . A plot of total charge passed, Q vs. square root of time, $t^{1/2}$ transforms the data into a linear relationship whose slope is $2nFACD^{1/2} \pi^{-1/2}$. The plot of Q vs. $t^{1/2}$ for $1.6 \times 10^{-7} \text{ M/cm}^3$ concentration of RO 12 is obtained. Using the value of the forward slope in the Cottrell equation, the diffusion coefficient 'D' of the compound is calculated as $2.261 \times 10^{-5} \text{ cm}^2\text{s}^{-1}$ at pH 4.0.

Controlled Potential Coulometry

Controlled potential coulometry was performed at pH 4.0 in the same cell setup. The coulometric 'n' value was determined after exhaustive electrolysis 10 ml of $3.2 \times 10^{-7} \text{ M dm}^{-3}$ RO 12 solution (pH 4.0) and it was found to be 2. This indicates the two-electron transfer in the electrooxidation of RO 12.

Reaction Mechanism

On the bases of above discussion the following mechanism was derived. The reactive orange 12 contain primary and secondary amine. The oxidation of primary amine takes place with 2 electron and proton reaction mechanism and the product is hydroxyl amine (Figure 3).

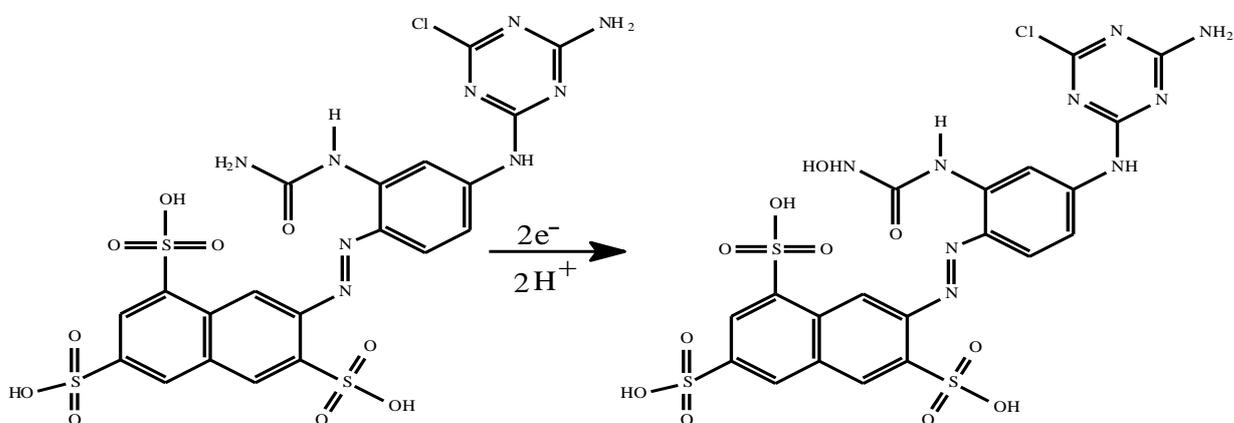


Fig 3, Reaction mechanism of reactive orange 12

DIFFERENTIAL PULSE STRIPPING VOLTAMMETRY

Having chosen the best pH trial experiments on GCE were carried out to optimize various parameters with the solution containing 300 ppb RO12. The adsorptive stripping voltammetry of RO12 was carried out on plain GCE at a positive potential and the stripping peak potential was close to the anodic oxidation potential observed in CV. This suggests the adsorptive accumulation and anodic stripping of the substrate. The factors affecting the electrode process were optimized and discussed as follows. Generally adsorptive stripping voltammetry involves two major steps. The analyte molecules were first accumulated by adsorption on the electrode surface from the bulk and then the accumulated molecules were stripped out giving specific signals. Hence, optimization of accumulation potential was done in the first part by fixing other parameters as in default setup. The effect of accumulation potential was studied by varying it from -800 to 800 mV. Maximum peak current was observed at 400 mV and it was fixed as the maximum peak current accumulation potential. The anion of the dye adsorbed easily at a positive potential due to electrostatic attraction. The effect of deposition time on the stripping signal in the range 15 to 90 sec was studied and the peak current reached maximum value at 60 seconds. The initial scan potential is another important parameter as it confirms the non-faradaic nature of the preconcentration step. It also controls both the peak potential and peak current in the stripping voltammogram. The influence of initial potential on the peak current was studied by varying from -1300 to -600 mV. Better response was observed at -1000 mV. The pulse height was varied from 25 to 200 mV and the maximum value was obtained at 150 mV. The effect of pulse width was studied in between 25 to 200 msec and was found that maximum peak current was at 75 msec. Effect of scan increment was studied by varying it from 2 to 20 mV. The peak current was linearly dependent on the scan increment up to 16 mV. The maximum peak current was observed at the scan increment 16 mV and hence it was chosen as the best and optimized scan increment. After fixing the above parameters the peak current response on scan rate was also studied. The scan rate was varied from 20 to 120 mV/s. Highly resolved stripping response with maximum peak current was obtained at the scan rate 50 mV/s. Hence, it was taken as the optimum scan rate for studying the effect of the concentration.

ANALYTICAL CHARACTERISTICS

The dependence of peak current on concentration was studied under the above fixed optimum parameters. The studies were carried out on glassy carbon electrode. A representative voltammogram is presented in figure 4. Experimental results showed that the peak current increased with increase in concentration of RO 12 on electrode and calibration plot arrived and are given figure 5. The reproducibility of the stripping signal was realized in terms of relative standard deviation for ten

identical measurements carried out at a concentration level of 300 ppb and the RSD values were 3.0% and the lower level of detection is 100 ppb.

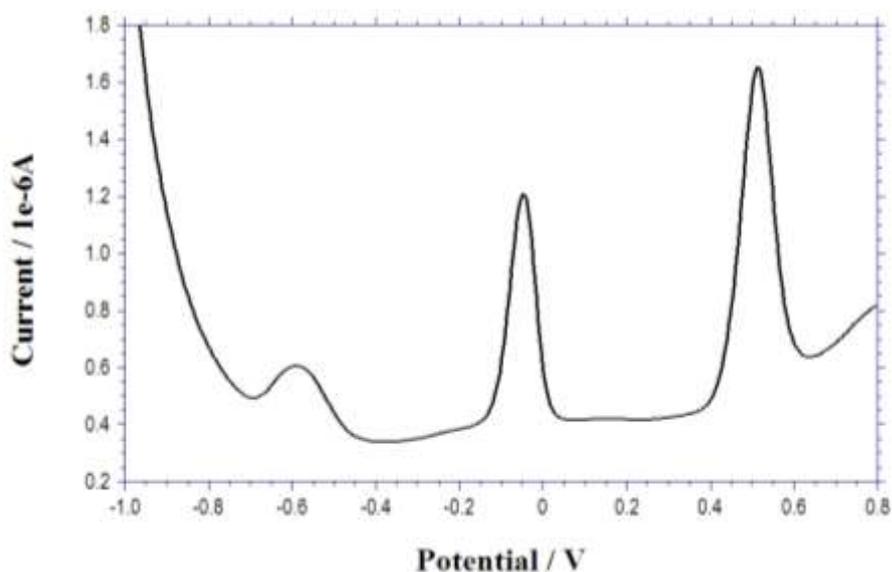


Figure 4, Differential pulse stripping voltammetry behavior of 300 ppb of RO 12 on GCE

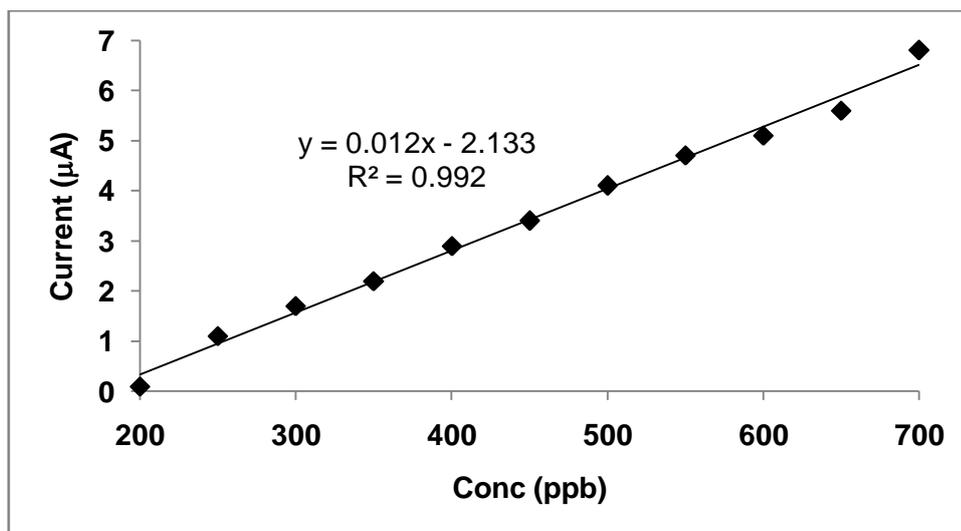


Figure 5, calibration plot of differential pulse stripping voltammetry

CONCLUSION

First attempt for studies of reactive orange 12 dye on glassy carbon electrode. The electrochemical behavior of RO12 was studied in different pH media. All the pH s shows three oxidation and two reduction peaks on GCE. Among the 13 medium of pH the cyclic voltammetric of RO12 illustrate best responds in pH 4.0. The best medium used to studies of further electrochemical behaviors. From five redox peaks the III anodic oxidation shows higher peak current and used for

other correlations. The peak current against scan rate and square root of scan rate and log of current responds vs log of scan rate revealed the nature of electrode reaction as irreversible diffusion controlled. The plot of Q vs. $t^{1/2}$ for 1.6×10^{-7} M/cm³ concentration of RO 12 is obtained 2.261×10^{-5} cm²sec⁻¹ diffusion coefficient 'D'. The bolometric 'n' value was determined after exhaustive electrolysis 10 ml of 3.2×10^{-7} M dm⁻³ RO 12 solution (pH 4.0) and it was found to be 2. The oxidation of primary amine takes place with 2 electron and proton reaction mechanism and the product is hydroxyl amine. The dependence of peak current on concentration was studied under the fixed optimum parameters. The amount of RO 12 was determined through differential pulse voltammetrically and the lower level of detection is 100 ppb.

REFERENCE

1. Mohan N, Balasubramanian N. In situ electrocatalytic oxidation of acid violet 12 dye effluent. *J Hazard Mater.* 2006; 136: 239–243.
2. Siddique M, Farooq R, Khan ZM, Khan Z, Shaukat SF. Enhanced decomposition of reactive blue 19 dye in ultrasound assisted electrochemical reactor. *Ultra Sonochem.* 2011; 18: 190–196.
3. Gouvea CA, Wypych F, Moraes SG, Duran N, Nagata N, Peralta-Zamora P. Semiconductor-assisted photocatalytic degradation of reactive dyes in aqueous solution. *Chemosphere.* 2000; 40(4): 433–440
4. Tang C, Chen V. The photocatalytic degradation of reactive black 5 using TiO₂/UV in an annular photoreactor. *Water Res.* 2004; 38(11): 2775–2781
5. Lucas M, Mertens V, Corbisier A, Vanhulle S. Synthetic dyes decolourisation by white rot fungi: development of original microtitre plate method and screening. *Enzym Microb Technol.* 2008; 42: 97–106
6. Rajkumar K, Muthukumar M. Response surface optimization of electro-oxidation process for the treatment of C.I. Reactive Yellow 186 dye: reaction pathways. *Appl Water.* 2015; doi: 10.1007/s13201-015-0276-0
7. Rueffer M, Bejan D, Bunce NJ. Graphite: an active or an inactive anode. *Electrochim Acta.* 2011; 56: 2246–2253
8. Rajkumar K, Muthukumar M, Mangalaraja RV. Electrochemical degradation of C.I. reactive orange 107 using gadolinium (Gd³⁺), neodymium (Nd³⁺) and samarium (Sm³⁺) doped cerium oxide nanoparticles, *International Journal of Industrial Chemistry.* 2015; 6: 285–295

9. Ajmal A, Rao RAK, Anwar S, Ahmad J, Ahmad R. Adsorption studies on rice husk: removal and recovery of Cd (II) from wastewater. *Bioresour. Technol.* 2003; 86: 147-149.
 10. Abd El-Latif MM, Ibrahim AM, El-Kady MF. Adsorption equilibrium, kinetics and thermodynamics of methylene blue from aqueous solutions using biopolymer oak sawdust composite. *J. Am. Sci.* 2010; 6: 276-283.
 11. Ozacar M, Sengil IA. Adsorption of metal complex dyes from aqueous solutions by pine sawdust. *Bioresour Technol.* 2005; 96: 791-795.
 12. Dizge N, Aydiner C, Demirbas E, Kobya M, Kara S. Adsorption of reactive dyes from aqueous solutions by fly ash: Kinetic and equilibrium studies. *J. Hazard. Mater.* 2008; 150: 737-747.
 13. Ghaedi AM, Vafaei A, Hafezi S, Mohagheghian M. Artificial neural network, equilibrium, kinetics and thermodynamics modeling of reactive orange 12 dye using rice husk. *Asian J. Chem.* 2013; 25: 817-826
 14. Dong Y M, He K, Zhao B, Yin Y, Yin L, Zhang A. Catalytic ozonation of azo dye active brilliant red X-3B in water with natural mineral brucite. *Catal. Commun.* 2007; 8: 1599-1603.
 15. Saratale R G, Saratale G D, Chang J S, Govindwar S P. Ecofriendly degradation of sulfonated diazo dye C.I. reactive green 19A using *micrococcus glutamicus* NCIM-2168. *Bioresour. Technol.* 2009; 100: 3897-3905.
 16. Ozdemir C, Oden MK, Sahinkaya S, Kalipci E. Color removal from synthetic textile wastewater by sono-fenton process. *Clean Soil Air Water*, 2011; 39(1): 60-67.
 17. Chung KT, Stevens S E J. Degradation of azo dyes by environmental microorganisms and helminthes. *Environ Toxicol Chem.* 1993; 12: 2121-2132.
 18. Serkan Sahinkaya, Decolorization of reactive orange 16 via ferrate(VI) oxidation assisted by sonication, *Turk J Chem.* 2017; 41: 577 – 586
 19. Iscen CF, Kiran I, Ilhan S. Biosorption of reactive black 5 dye by *penicillium restrictum*: the kinetic study. *Journal of Hazardous Materials.* 2007; 143: 335–340.
 20. Osma JF, Saravia V, Toca-Herrera JL, Couto SR. Sunflower seed shells: A novel and effective low-cost adsorbent for the removal of the diazo dye Reactive Black 5 from aqueous solutions. *Journal of Hazardous Materials.* 2007; 147: 900–905.
 21. Maytham Kadhim Obaid, Luqman Chuah Abdullah, Intidhar Jabir Idan. Removal of reactive orange 16 dye from aqueous solution by using modified kenaf core fiber, *Journal of Chemistry*, 2016, Article ID 4262578, 7 pages, <http://dx.doi.org/10.1155/2016/4262578>.
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